

Reproductive dynamics of Diplostraca in the Black Sea: insights from Hamsilos Bay, Sinop, Türkiye

Funda ÜSTÜN^{*} 

Department of Hydrobiology, Faculty of Fisheries, Sinop University, Sinop, Türkiye

Received: 14.02.2025 • Accepted/Published Online: 18.05.2025 • Final Version: 17.07.2025

Abstract: In the coastal marine ecosystem of the Black Sea, Diplostraca is the second most dominant mesozooplankton after Copepoda. Its parthenogenetic reproductive strategy enables significant population growth from spring (March) to fall (December), contributing substantially to mesozooplankton abundance. This study investigated the fecundity, embryonic development stages, and population structure of four of Diplostraca species in Hamsilos Bay: *Evadne spinifera*, *Pleopis polyphemoides*, *Penilia avirostris*, and *Pseudevadne tergestina*. Samples were taken every month from three stations between July 2015 and June 2016. Population analysis showed that parthenogenetic females were dominant, reaching up to 100% of the population. Gamogenetic females, females without embryos, and males constituted 23%, 33%, and 9% of the population, respectively. Embryonic development stages were species-specific: stage A embryos were predominant in *Ps. tergestina*, *Pl. polyphemoides*, and *E. spinifera*, whereas stage C embryos were most common in *Pe. avirostris*. The ultimate stage, stage D, was rare across all species. The number of embryos per parthenogenetic female varied from 1 to 5 in *E. spinifera*, 2–8 in *Pe. avirostris*, 1–8 in *Pl. polyphemoides*, and 1–6 in *Ps. tergestina*. Diplostraca primarily reproduced via parthenogenesis, enhancing their contribution to the mesozooplankton community in Hamsilos Bay. The number of embryos observed was lower compared to a similar study performed in the Marmara Sea, possibly due to distinct regional environmental differences such as nutrient availability and temperature. This study bridges a critical gap in understanding the reproductive dynamics of Diplostraca in semiencloded seas and factors affecting them, establishing a valuable reference for future research.

Key words: Diplostraca, reproductive biology, population structure, embryo number, embryo stage

1. Introduction

Diplostraca Gerstaecker, 1866, is a superorder of marine zooplankton. They are a vital component of the marine zooplankton community, playing a key role in the marine planktonic food web. Diplostraca are primary consumers of phytoplankton (He et al., 2023) and a major food source for zooplankton (Schroeder et al., 2022; Uye and Liang, 2022), fish larvae (Sabatés et al., 2015), and planktivorous fish (Lankov et al., 2010). Despite their global distribution in oceans, Diplostraca show relatively low species diversity in marine environments compared to freshwater environments, the latter including more than 620 species (Forró et al., 2008). To date, only eight Diplostraca species have been identified in marine habitats (Onbé, 1999).

The reproductive cycle of Diplostraca alternates between two stages: asexual reproduction (parthenogenesis) and sexual reproduction (gamogenesis) (Egloff et al., 1997). During the parthenogenetic stage, genetically identical embryos (Hebert and Ward, 1972) develop in the brooding chamber beneath the carapace of parthenogenetic females and are discharged into the sea as free-swimming

adult individuals (Platt and Yamamura, 1986). When environmental conditions become unfavorable for population growth, Diplostraca transition to gamogenesis (Egloff et al., 1997). In this stage, males emerge, and gamogenetic females produce genetically diverse resting eggs that settle on the seabed (Onbé, 1985; Egloff et al., 1997).

Diplostraca species often achieve high population densities during certain times of the year, particularly from early summer to late fall. This rapid increase, driven by parthenogenetic reproduction, significantly enhances zooplankton abundance and species diversity in coastal marine ecosystems. However, their population density declines just as rapidly, eventually disappearing from the planktonic community (Kodama et al., 2021).

While the composition, seasonal abundance peaks, distribution, and environmental relationships of Diplostraca species in Turkish seas are well-documented (Buyukates and İnanmaz, 2007, 2009; Isinibilir, 2009; Gülşahin and Tarkan, 2012; Toklu Alıçlı et al., 2014; Terbiyik Kurt and Polat, 2014; Terbiyik Kurt and Polat,

* Correspondence: fund austun@gmail.com

2017; Terbiyik Kurt et al., 2018; Üstün et al., 2016, 2018; Üstün, 2019; Killi, 2020; Yildiz, 2025), studies on their reproductive strategies and fecundity have been confined to the Sea of Marmara (Bedikoğlu et al., 2022). The growth and reproductive cycle of *Penilia avirostris* were previously investigated exclusively by Pavlova (1959) in Sevastopol Bay in the Black Sea. Although Pavlova's study (1959) provides an important foundation, it focused on a single species, and the considerable time elapsed since its publication underscores the need for updated and comprehensive data. To address this gap, the present study examines the sex ratios, embryo number, and embryonic development of *Evadne spinifera*, *Pleopis polyphemoides*, *Pe. avirostris*, and *Pseudevadne tergestina* in Hamsilos Bay in the Black Sea. By expanding on Pavlova's findings (1959), this study provides detailed and up-to-date insights into the reproductive dynamics of Diplostraca in the region. This study serves as a valuable foundation for future studies, as it is the first to explore Diplostraca reproductive biology in the Turkish waters of the Black Sea.

2. Materials and methods

Sinop is situated on the Boztepe Peninsula on the northernmost coast of Anatolia, which extends into the Black Sea. This peninsula features two natural harbors oriented to the east and west, known as Büyük Liman (inner harbor) in the southeast and Akliman (outer harbor) in the northwest. Hamsilos Nature Park was designated a natural protected area in 2007¹ and is an ecologically significant region requiring conservation (Öztekin et al., 2024). Hamsilos Bay is located within the Akliman region of the park, and is therefore exposed to storms originating from the northwest (Uygun and Üstün, 2024).

Sampling was performed on a monthly basis at three stations between July 2015 and June 2016 in Hamsilos Bay. The geographical coordinates of the stations were as follows: station 1 (42°03'52"N, 35°03'14"E), station 2 (42°04'5"N, 35°02'59"E), and station 3 (42°04'12"N, 35°02'45"E). All stations were situated near the mouth of the bay, where it connects to the open sea at a depth of 30 m (Figure 1).

Surface seawater temperature and salinity were recorded in situ at each station using a 6600 MDS multiparameter device (YSI Inc., Yellow Springs, OH, USA) during sampling surveys. Chlorophyll-a concentrations ($\mu\text{g L}^{-1}$) of surface seawater were determined using the acetone extraction method described by Parsons et al. (1984). The average values of surface seawater chlorophyll-a, salinity, and temperature from the three stations were calculated and used in subsequent analyses.

Sampling was conducted through daytime vertical hauls from the seafloor to the surface using a plankton net (0.2 m² mouth area and 112 μm mesh size). The sample was promptly fixed in a solution of formaldehyde, seawater, and borax (4% v/v), for subsequent quantitative and qualitative analyses.

The abundance of Diplostraca species was calculated as previously described by Postel et al. (2000). Diplostraca species were identified and counted under a stereomicroscope in two 1 mL subsamples taken from each main sample with a Henson-Stempel pipette that had a known volume of either 75 mL or 100 mL. The mean values obtained from the replicates were extrapolated to estimate the total abundance per sample. Abundance values were averaged across the three stations and expressed as the mean $\pm$ standard deviation of individuals per cubic meter (ind. m⁻³).

For each station, thirty individuals of each target species (*E. spinifera*, *Pe. avirostris*, *Ps. tergestina*, and *Pl. polyphemoides*) were randomly selected from the monthly samples. Sex and brood composition analyses were not conducted in months with low individual counts (<30 individuals per station), such as March and June 2016, due to insufficient sample size. The selected individuals were categorized based on their sex and whether they contained embryos or eggs in their brood pouches. For *E. spinifera*, *Pl. polyphemoides*, and *Ps. tergestina*, individuals were divided into four groups: female (F: female without embryos), parthenogenetic female (PF: female with embryos), gametogenetic female (GF: female with resting eggs) and male (M). For *Pe. avirostris*, an additional juvenile class (J: individuals not producing embryos, <500 μm) was included. All individuals were examined under a stereomicroscope, and the brood pouches of PF and GF individuals were carefully opened using a fine dissecting needle inserted through the anus.

The number of embryos in PF individuals was recorded, and their developmental stages were determined. Resting eggs (Re) extracted from GF individuals were also counted separately. Based on morphological characteristics, embryos from PF individuals were classified into four developmental stages: stage A (the first stage), stage B, stage C, and stage D (final stage) (Wong et al., 2004; Bedikoğlu et al., 2022).

Fecundity was defined in this study as the mean number of embryos per PF.

Spearman's rank correlation analysis was performed to assess the relationship between environmental variables (temperature, salinity, and chlorophyll-a) and the abundance of Diplostraca species. The analysis was conducted using R software version 4.2.3.²

¹Republic of Türkiye Ministry of Agriculture and Forestry (2023). Sinop-Hamsilos Natural Park (in Turkish). Website <https://bolge10.tarimorman.gov.tr/Menu/48/Sinop-Hamsilos-Tabiat-Parki> [accessed 03 March 2023].

²R Core Team (2023). The R Project for Statistical Computing [online]. Website <https://www.R-project.org> [accessed 14 April 2025].

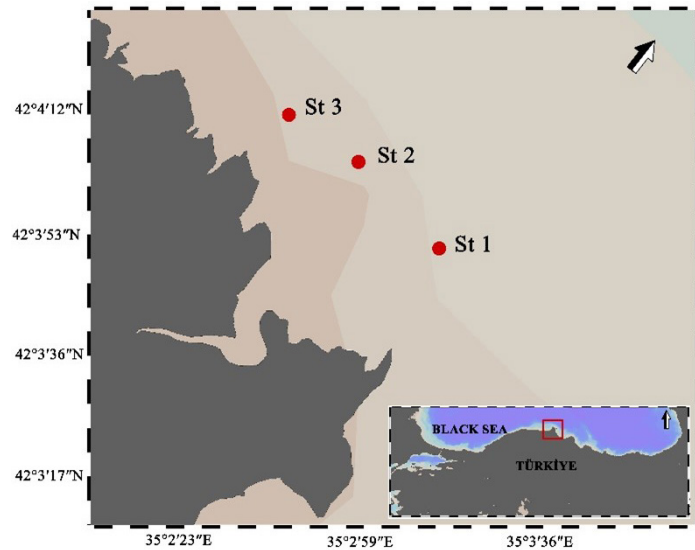


Figure 1. Map of the sampling area.

The readr package³ was used for data import, and the Hmisc package⁴ for correlation analysis.

3. Results

3.1. Environmental parameters, species composition, and seasonal occurrence of Diplostraca

Monthly variations in surface seawater temperature, salinity, chlorophyll-a concentrations, and Diplostraca abundance were observed in Hamsilos Bay. The minimum temperature, 8.5 ± 0.01 °C, was observed in February 2016, whereas the maximum temperature, 25.8 ± 0.18 °C, occurred in August 2015. Salinity fluctuated from 17.79 ± 0.08 in September 2015 to 18.96 ± 0.02 in May 2016. Chlorophyll-a concentration ranged from 0.14 ± 0.03 µg L⁻¹ in June 2016 to 0.87 ± 0.03 µg L⁻¹ in February 2016 (Figure 2).

The abundance of Diplostraca showed distinct seasonal trends, peaking in September 2015 with 1897.5 ± 200.53 ind. m⁻³ (Figure 2). Relatively high abundance values were also recorded in August 2015 (574.86 ± 220.87 ind. m⁻³) and July 2015 (217.5 ± 21.03 ind. m⁻³). Conversely, the lowest abundance values were recorded in March 2016 (0.28 ± 0.48 ind. m⁻³) and June 2016 (3.72 ± 4.14 ind. m⁻³).

The warm water species *Pe. avirostris*, *Ps. tergestina*, and *E. spinifera* first appeared in July 2015 (23.94 °C) and reached peak abundance in September 2015 (24 °C). These species were not observed in the samples after November (15.46 °C), October (19.58 °C), and September

2015, respectively. *Pl. polyphemoides* achieved maximum abundance in July 2015 under high-temperature conditions (23.94 °C), but it was not observed after November 2015. The species reemerged in March 2016 at lower temperatures (8.85 °C) (Figure 3).

The relationship between the abundances of Diplostraca species and environmental variables (temperature, salinity, and chlorophyll-a) is presented in Table 1. The abundance values of Diplostraca species showed a positive and statistically significant correlation with temperature ($p < 0.001$). This relationship was particularly strong for Diplostraca, *Pe. avirostris*, and *Ps. tergestina* ($r = 0.892$, $r = 0.832$, $r = 0.807$, respectively). In contrast, salinity showed a negative and statistically significant correlation for most species ($p < 0.001$), with Diplostraca ($r = -0.780$) and *Pe. avirostris* ($r = -0.767$) showing the strongest negative correlations. Correlations with chlorophyll-a were weaker; however, significant negative relationships were observed for some species, such as *E. spinifera* ($r = -0.508$, $p < 0.005$).

3.2. Population structure

PFs were the most abundant Diplostraca population in this study, comprising 74.4% to 100% of individuals. The population abundance increased rapidly from July 2015 to October 2015 due to parthenogenetic reproduction. GFs were observed more frequently than Ms, with the proportion of GF individuals varying by species. The proportion of GF individuals was 1.1% for *Pl. polyphemoides* and *Pe. avirostris* and 15.5% for *E. spinifera*. No GF individuals

³ Wickham H, Hester J, Francois R, Bryan J, Bearrows S, Posit Software, PBC (2022). readr: Read Rectangular Text Data [online]. Website <https://cran.r-project.org/package=readr> [accessed 14 April 2025].

⁴ Harrell FE Jr, Dupont C (2022). Hmisc: Harrell Miscellaneous [online]. Website <https://cran.r-project.org/package=Hmisc> [accessed 14 April 2025].

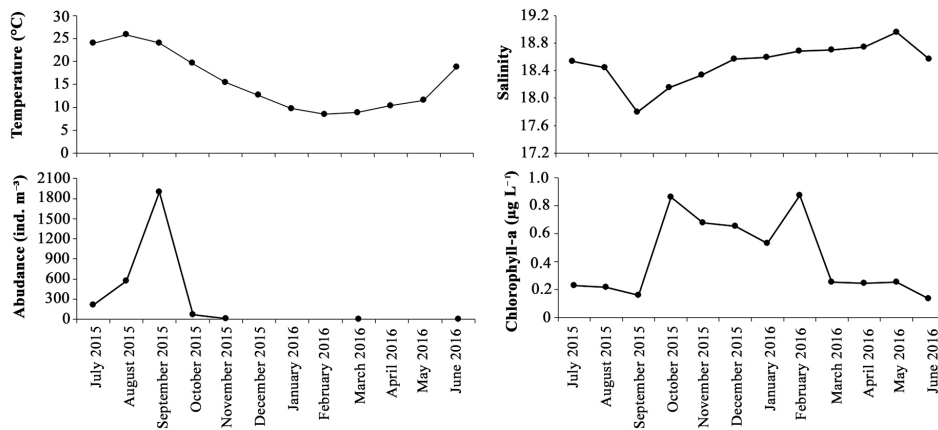


Figure 2. Monthly variation of mean surface seawater temperature (°C), salinity, chlorophyll-a ($\mu\text{g L}^{-1}$) values, and Diplostraca abundance (ind. m^{-3}) in Hamsilos Bay.

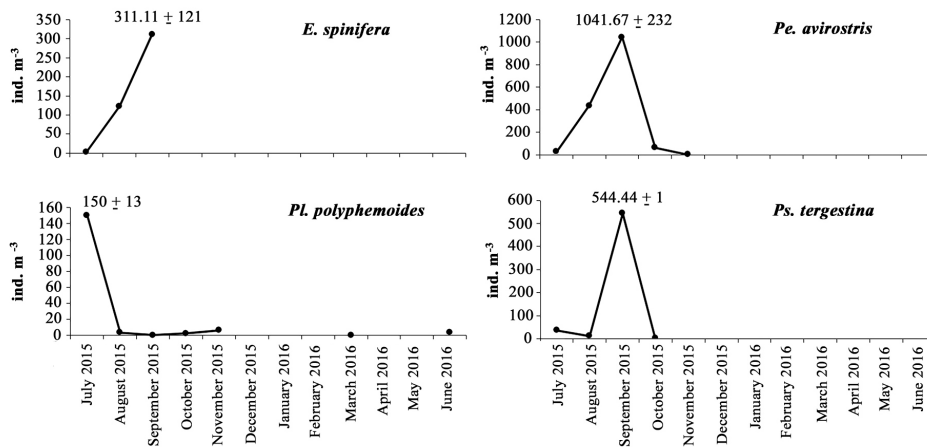


Figure 3. Monthly variations in the mean abundance of four Diplostraca species in Hamsilos Bay.

were detected in the *Ps. tergestina* population. In this study, the proportion of Ms in the population was around 3%, with no *Pl. polyphemoides* and *Pe. avirostris* males found. The proportion of F in the Diplostraca population ranged from 3.1% to 22.2%. Juvenile individuals were identified only in the *Pe. avirostris* species, ranging from 2.2% to 7.8%. This suggests that the *Pe. avirostris* population has an active reproductive cycle (Figure 4).

All species were present at all stations. Except for *Pl. polyphemoides* in August 2015 (30 individuals from only one station), all species—including *Pl. polyphemoides* in July 2015—were present at all three stations, with 30 individuals examined per station. PFs were notably more numerous during the summer months across all species. GFs, although observed in lower numbers, showed a notable increase in *E. spinifera* in September 2015. Additionally, only one resting egg was observed per GF individual across all species (Table 2).

A higher abundance of embryos at the initial developmental stage (stage A) was detected in *E. spinifera* (46% to 48%), *Ps. tergestina* (38% to 51%), and *Pl. polyphemoides* (40% to 43%). On the other hand, stage C was most frequently noted in *Pe. avirostris* (48% to 68%). Stage D, the last phase of embryonic development, was the least common stage observed among all species (1% to 5%). In every species analyzed, the mean number of embryos in stage A was higher compared to the other embryonic stages (Figure 5).

3.3. Reproductive strategies

The embryo number of PFs varied across the four Diplostraca species studied: *Pe. avirostris* carried between two and eight embryos, *E. spinifera* between one and five, *Ps. tergestina* between one and six, and *Pl. polyphemoides* between one and eight. The highest mean embryo number (5.77 ± 0.80 embryos PF^{-1}) was observed in *Pe. avirostris*

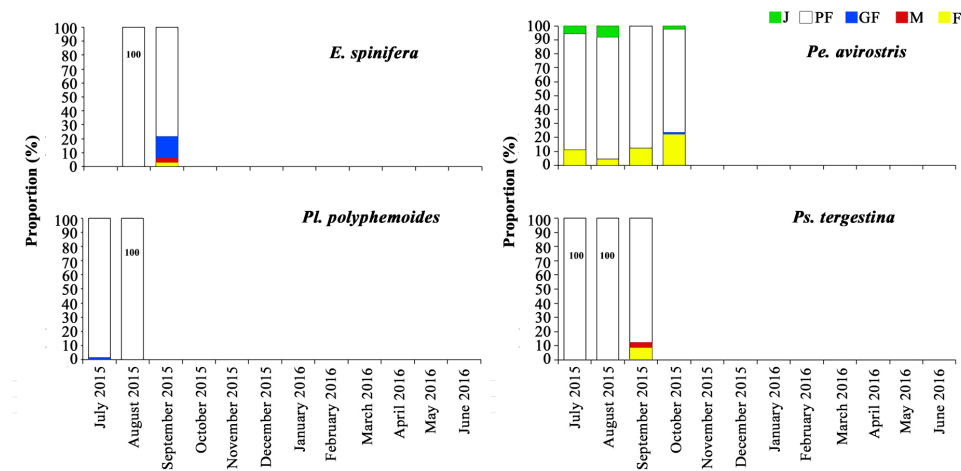
Table 1. Spearman correlation coefficients between environmental variables and the abundance of Diplostraca species.

Taxa	Temperature	Salinity	Chlorophyll-a
Diplostraca	0.892***	-0.780***	-0.364*
<i>E. spinifera</i>	0.705***	-0.460*	-0.508**
<i>Pe. avirostris</i>	0.832***	-0.767***	-0.177
<i>Pe. polyphemoides</i>	0.655***	-0.485*	-0.198
<i>Ps. tergestina</i>	0.807***	-0.644***	-0.292

Significance levels: * $p < 0.05$, ** $p < 0.005$, *** $p < 0.001$.

Table 2. Numbers of PF, GF, and Re for each species based on the combined data of individuals collected from three stations each month (30 individuals per station for each species).

Collection date	Species	PF	GF	Re
July 2015	<i>E. spinifera</i>	-	-	-
	<i>Pe. avirostris</i>	75	-	-
	<i>Pl. polyphemoides</i>	89	1	1
	<i>Ps. tergestina</i>	90	-	-
August 2015	<i>E. spinifera</i>	90	-	-
	<i>Pe. avirostris</i>	79	-	-
	<i>Pl. polyphemoides</i>	30	-	-
	<i>Ps. tergestina</i>	90	-	-
September 2015	<i>E. spinifera</i>	73	14	14
	<i>Pe. avirostris</i>	79	-	-
	<i>Pl. polyphemoides</i>	-	-	-
	<i>Ps. tergestina</i>	79	-	-
October 2015	<i>E. spinifera</i>	-	-	-
	<i>Pe. avirostris</i>	67	1	1
	<i>Pl. polyphemoides</i>	-	-	-
	<i>Ps. tergestina</i>	-	-	-


Figure 4. Monthly variation in the population stage composition of four Diplostraca species in Hamsilos Bay, based on combined data from three stations.

in July 2015, under higher chlorophyll-a concentrations ($0.86 \mu\text{g L}^{-1}$) and elevated temperatures (25.8°C). High reproductive success was recorded for *Pe. avirostris* in July, September, and October 2015, whereas *Pl. polyphemoides* showed high reproductive success in August 2015 (Table 3).

4. Discussion

4.1. Seasonal occurrence of Diplostraca

Diplostraca plays a crucial role in the mesozooplankton community of coastal marine ecosystems, particularly during warm seasons (Du et al., 2022). Their reproductive strategies are key to population dynamics, as larval

fish depend on zooplankton such as copepods and diplostraceans for feeding (Sabatés et al., 2015). Studies suggest that larval fish may preferentially target diplostraceans over copepods, with those primarily feeding on diplostraceans showing higher survival rates compared to individuals consuming other zooplankton (Mayer and Wahl, 1997; Graeb et al., 2004). Through parthenogenetic reproduction, Diplostraca species can rapidly achieve high population densities, providing a vital food source for fish larvae, especially during critical developmental stages. This may increase survival rates and enhance recruitment success in the early ontogenetic stages of fish (Uygun and Hoşsucu, 2024). Thus, fluctuations in Diplostraca

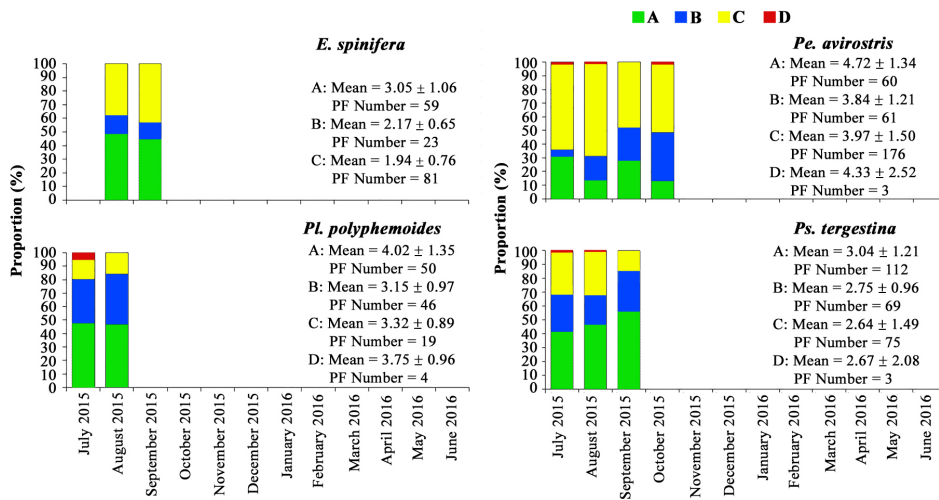


Figure 5. Monthly variation of embryo stages of PFs of four Diplostraca species in Hamsilos Bay, based on the combined data from three stations.

Table 3. Fecundity of PFs of four Diplostraca species in Hamsilos Bay.

Species	Number of embryos	Number of embryos PF ⁻¹					Temperature	Salinity	Chlorophyll-a
		July 2015	August 2015	September 2015	October 2015	Overall Mean $\pm$ SD			
<i>E. spinifera</i>	1–5		2.38 ± 1.06	2.37 ± 0.95		2.37 ± 1.01	24 (September)–25.8 (August)	17.8 (September)–18.4 (August)	0.16 (September)–0.22 (August)
<i>Pe. avirostris</i>	2–8	5.77 ± 0.80	2.41 ± 0.59	4.04 ± 0.82	4.28 ± 1.00	4.1 ± 1.45	19.6 (October)–25.8 (August)	17.8 (September)–18.5 (July)	0.16 (September)–0.86 (October)
<i>Pl. polyphemoides</i>	1–7	3.27 ± 1.00	4.43 ± 1.30			3.56 ± 1.19	23.9 (July)–25.8 (August)	18.4 (August)–18.5 (July)	0.22 (August)–0.23 (July)
<i>Ps. tergestina</i>	1–7	3.70 ± 1.20	2.52 ± 1.11	2.23 ± 0.91		2.84 ± 1.26	23.9 (July)–25.8 (August)	17.8 (September)–18.5 (July)	0.16 (September)–0.23 (July)

abundance, driven by seasonal and environmental factors, can significantly impact the growth of fish larvae and recruitment success within the Black Sea ecosystem. Investigating these dynamics is necessary for predicting the function of mesozooplankton communities in coastal food webs and evaluating overall ecosystem productivity.

Environmental factors, particularly temperature, significantly affect the seasonal distribution and abundance of Diplostraca (Pavlova, 1959; Marazzo and Valentin, 2003a; Yildiz, 2025). Diplostraca species emerge in early spring, with maximum abundance observed in summer and autumn months. They become scarce or absent in winter (Isinibilir, 2009; Terbiyik Kurt and Polat, 2014), with their resting eggs remaining on the seabed (Ramírez and Pérez Seijas, 1985). The abundance of Diplostraca species increased throughout the summer and autumn periods in this research. This seasonal distribution aligns with the ecological characteristics of the species and supports findings from similar studies conducted in the Black Sea (Lebedeva et al., 2015; Üstün et al., 2016, 2018; Bişinicu et al., 2023; Yildiz, 2025).

4.2. Population structure

Gametogenetic reproduction, one of the two reproductive strategies of Diplostraca, is essential for ensuring the continuity of their lineage. Within the population structures of these species, GF and M represent individuals using this reproductive strategy. In general, gametogenetic individuals constitute only a small proportion of the population in coastal marine regions (Hairston and Cáceres, 1996; Miyashita et al., 2010). However, there are exceptions, as observed in this study for *E. spinifera*. For instance, gametogenetic individuals of *Ps. tergestina* make up 13.6% of the population in the Atlantic Ocean (Oghenekaro and Chigbu, 2019), and between 3% and 7% in the Gulf of Mexico (Cházaro-Olvera et al., 2024). In some cases, this proportion exceeds 70% in the South Atlantic Ocean (Marazzo and Valentin, 2004a). The proportion ranges from 20% to 50% in the South Atlantic Ocean for *Pl. polyphemoides* (Marazzo and Valentin, 2003b). Marazzo and Valentin (2003a) suggested that high population densities could increase the probability of encounters between female and male specimens, thus enhancing the proportion of gametogenetic individuals. In our study, the high proportion of gametogenetic individuals observed in *E. spinifera* suggests that this species may use an alternative reproductive strategy in response to environmental conditions. Furthermore, the preference for gametogenetic reproduction may serve as a strategy to enhance genetic diversity and increase the adaptive capacity of the species.

Although the mechanisms underlying the transition from parthenogenetic to gametogenetic reproduction in marine Diplostraca remain unclear (Marazzo and Valentin, 2003a), GF and M individuals in marine Diplostraca species typically appear during periods of high

population density and reach their maximum abundance just before the population disappears from the plankton (Egloff et al., 1997). GF and M individuals were observed in *E. spinifera*, *Pl. polyphemoides*, and *Ps. tergestina* during the months when their abundance peaked (September 2015 and July 2015) in the present study. However, GF *Pe. avirostris* individuals were detected in October 2015, after the peak abundance in September. A similar trend has been documented for *Pe. avirostris* in prior investigations from the Western Mediterranean Sea (Atienza et al., 2008), the Sea of Marmara (Bedikoğlu et al., 2022), the Gulf of Mexico (Cházaro-Olvera et al., 2024), and the South Atlantic Ocean (Marazzo and Valentin, 2003a, 2004b). Gametogenetic individuals are less frequently observed due to insufficient sampling frequency and the analysis of only subsamples, making their detection more challenging. According to Egloff et al. (1997), different Diplostraca species may exhibit varying responses to the same environmental cues within the same time and region, with gametogenetic reproduction occurring in some species but not in others under similar conditions.

In Diplostraca species, individuals in the last stage (stage D) of embryonic development have pigmented eyes (Platt and Yamamura, 1986) and are more easily detected by planktivorous fish and other predators, such as chaetognaths (Mullin and Onbé, 1992; Wong et al., 2008). Therefore, PFs carrying developed embryos are generally found in the water column at night to avoid intense predation pressure (Mullin and Onbé, 1992). This study was carried on during the daytime, which might explain the low number of stage D embryos observed.

4.3. Reproductive strategies

Few studies have characterized the reproductive biology of Diplostraca species, including *E. spinifera*, *Pl. polyphemoides*, *Pe. avirostris*, and *Ps. tergestina*. The limited detailed research on these species has been conducted in five regions: the South Atlantic Ocean, the Western Mediterranean Sea, the China Sea, the Gulf of Mexico, the Inland Sea of Japan, and the Sea of Marmara.

The varying embryo numbers for *Pe. avirostris* in Hamsilos Bay is generally consistent with the values reported for the Inland Sea of Japan (Mullin and Onbe, 1992), the Catalan Sea (Atienza et al., 2008), and Gulf of Mexico (Cházaro-Olvera et al., 2024). However, the values noted in the present study were lower than those documented for the Gulf of Mexico (Della Croce and Angelino, 1987), the Southern China Sea (Tang et al., 1995), the coastal region of Brazil (Marazzo and Valentin, 2003a, 2004b; Miyashita et al., 2010), and the Sea of Marmara (Bedikoğlu et al., 2022). The embryonic fecundity observed in this study is similar to the values documented for the coastal regions of the Inland Sea of Japan (Mullin and Onbe, 1992), the Southern China Sea (Tang et al., 1995), the coastal region of Brazil (Marazzo and Valentin, 2004b; Miyashita et al., 2010), and the Gulf of Mexico (Cházaro-Olvera et al., 2024) (Tables 3 and 4).

Table 4. Fecundity of PFs of four Diplostetraca species from different environments.

Species	Number embryos	Number of embryos PF ⁻¹	Study area	References	Temperature	Salinity
<i>Pe. avirostris</i>	1-11		Black Sea	Pavlova (1959)		
	1-10		Gulf of Mexico	Della Croce and Angelino (1987)	27.0-30.9	34.7-38.0
	1-8	3.25	Inland Sea of Japan	Mullin and Onbé (1992)	24-26	
	1-14	4.4	South China Sea	Tang et al. (1995)	16-32	7.3-37.2
	3-9		Brazil – South Atlantic Ocean	Marazzo and Valentin (2003a)	20.5-27.5	23-26
	1-11	4.35	Brazil – South Atlantic Ocean	Marazzo and Valentin (2004b)	21.2-28.5	24.6-35.6
	1-8	3.21 ± 0.09 in 2003 4.23 ± 0.99 in 2004	Spain – Northwest Mediterranean	Atienza et al. (2008)	16-27	
	1-12	3.6 ± 1.5	Brazil – South Atlantic Ocean	Miyashita et al. (2010)	14.8-28.2	31.1-36
	1-11		Sea of Marmara	Bedikoglu et al. (2022)	11.5-24.7	21.8-28.8
	2-7	3.7 ± 0.26	Gulf of Mexico	Cházaro-Olvera et al. (2024)	26.5-29.3	33.4-35.9
<i>Ps. tergestina</i>	3-12	7.76	Chesapeake Bay, USA	Bryan (1979)	25.2-29.6	14.4-31.2
	1-8		Gulf of Mexico	Della Croce and Angelino (1987)	16.4-30.9	29.92-38.03
	1-7	3.3	Inland Sea of Japan	Mullin and Onbé (1992)	24-26	
	2-6	3.3	Gulf of Mexico	Mullin and Onbé (1992)	27-28	
	1-14	4.6	South China Sea	Tang et al. (1995)	16-32	7.3-37.2
	2-8	4.3 ± 0.16	Gulf of Mexico	Cházaro-Olvera et al. (2024)	26.5-29.3	33.4-35.9
	4-8	6.36 ± 0.77	Brazil – South Atlantic Ocean	Marazzo and Valentin (2004a)	21-27	
	1-14	6.49	Brazil – South Atlantic Ocean	Marazzo and Valentin (2004b)	21.2-28.5	24.7-35.6
	1-10	3.7 ± 1.5	Brazil – South Atlantic Ocean	Miyashita et al. (2011)	14.8-28.2	31.1-35.7
	1-19	6.57 ± 0.15	Atlantic Ocean	Oghenekaro and Chigbu (2019)	20.6-23.7	30-35.5
<i>Pl. polyphemoides</i>	1-9		Sea of Marmara	Bedikoglu et al. (2022)	10.4-24.2	21.9-27.3
	5-8	6.71 ± 0.78	Brazil – South Atlantic Ocean	Marazzo and Valentin (2003b)	21-28	23-35
	1-10	3.6 ± 0.3	Baltic Sea	Pöllupüü et al. (2010)	5-25	1-6
	1-6	2.9 ± 1.5	Brazil – South Atlantic Ocean	Miyashita et al. (2011)	15.3-25	33.7-35.7
	1-13		Sea of Marmara	Bedikoglu et al. (2022)	6.9-24.5	21.8-28.5
<i>E. spinifera</i>	1-10	5 ± 1.5	Brazil – South Atlantic Ocean	Miyashita et al. (2011)	14.8-27.2	31.1-35.6
	4-6	4.3 ± 0.33	Atlantic Ocean	Oghenekaro and Chigbu (2019)		

The embryo number range for *Ps. tergestina* was similar to that observed in the Inland Sea of Japan (Mullin and Onbé, 1992) and the Gulf of Mexico (Della Croce and Angelino, 1987; Cházaro-Olvera et al., 2024). However, embryo numbers reported for Chesapeake Bay (Bryan, 1979), the Southern China Sea (Tang et al., 1995), the coastal region of Brazil (Marazzo and Valentin, 2004a, 2004b; Miyashita et al., 2011), the Atlantic Ocean (Oghenekaro and Chigbu, 2019), and the Sea of Marmara (Bedikoğlu et al., 2022) were found to be quite high and varied widely. The embryonic fecundity for *Ps. tergestina* in the present study is significantly lower than those reported for the Chesapeake Bay (Bryan, 1979), the Inland Sea of Japan (Mullin and Onbé, 1992), the Gulf of Mexico (Della Croce and Angelino, 1987; Mullin and Onbé, 1992; Cházaro-Olvera et al., 2024), the coastal region of Brazil (Marazzo and Valentin, 2004a, 2004b; Miyashita et al., 2011), and the Atlantic Ocean (Oghenekaro and Chigbu, 2019) (Tables 3 and 4).

The embryo number range and the mean number of embryos for *Pl. polyphemoides* observed in the current study were similar to those reported from the coastal region of Brazil (Miyashita et al., 2011), but lower than the higher embryo numbers reported in the Sea of Marmara and the Baltic Sea (Pöllupüü et al., 2010; Bedikoğlu et al., 2022) (Tables 3 and 4). Compared to the Sea of Marmara and the Baltic Sea, the lower number of embryos observed in this study may be associated with the relatively higher seawater temperature. In the Sea of Marmara, the highest embryo number for this species was recorded at 10.7 °C (Bedikoğlu et al., 2022), while studies in the Baltic Sea reported a negative correlation between parthenogenetic fecundity and seawater temperature (Pöllupüü et al., 2010). In both regions, higher embryo production by PFs was observed at lower temperatures.

Similarly, the embryo number range and the mean number of embryos of *E. spinifera* determined in this study were notably lower than those reported the South Atlantic Ocean (Miyashita et al., 2011) and the Atlantic Ocean (Oghenekaro and Chigbu, 2019) (Tables 3 and 4). In the South Atlantic Ocean, a negative correlation between seawater temperature and embryo number was identified. Moreover, in the Atlantic Ocean, *E. spinifera* was observed only in July 2012, coinciding with a peak in chlorophyll-a concentration ($5.8 \mu\text{g L}^{-1}$), suggesting that food availability may play a critical role in reproductive output. Accordingly, the relatively high seawater temperatures and low chlorophyll-a concentrations recorded in Hamsilos Bay during the present study may have contributed to the lower embryo numbers observed.

Fecundity values reported from various marine regions such as the South Atlantic Ocean, the Gulf of Mexico, the Sea of Marmara, and the Black Sea (this study) vary

(Tables 3 and 4). These differences suggest geographical variation in reproductive capacity that may be influenced by regional environmental and biological factors. Changes in environmental conditions and population density are known to affect the reproductive dynamics of Diplostraca species (Egloff et al., 1997; Cházaro-Olvera et al., 2024). Such variation in reproductive capacity and embryo numbers may be associated with differences in nutrient levels, salinity, temperature, regional ecological conditions, methodological approaches, and population genetics.

5. Conclusion

The Diplostraca species in Hamsilos Bay successfully reproduce during the short warm period in summer and early autumn, making significant contributions to the mesozooplankton community. Our findings showed that these species predominantly reproduce through parthenogenesis during their presence in the water column. Embryo numbers were lower compared to similar species in the Sea of Marmara, which may be attributed to environmental and genetic factors. The study area is free from pollution and maintains its natural structure, further enhancing the reliability of the surveyed data. Achieving a more comprehensive understanding of the predation pressures, ecological roles, and population dynamics of Diplostraca in the Black Sea calls for long-term studies with frequent and regular sampling are recommended. As a pioneering investigation into the reproductive strategies of Diplostraca species inhabiting the Turkish waters of the Black Sea, this research serves as an important reference for understanding the reproductive patterns of species in the region.

Acknowledgment

Abundance data, along with temperature, salinity, and chlorophyll-a measurements from station 2 were obtained from Üstün (2019). In this study, the data from station 2 were combined with data from two additional stations and are presented as averages. This study was delivered as an oral presentation at the VI. International Agricultural, Biological and Life Science Conference held in Edirne, Türkiye, September 18–20, 2024.

The author would like to thank the crew of RV Zıpın for their support during the sampling, Dr. Dalida Bedikoğlu for her assistance in identifying the embryos, and Dr. Orçin Uygün for preparing the sampling area map and conducting the statistical analysis.

This study was supported by the Sinop University Scientific Research Project (SÜF-1901–14-04).

Conflict of interest

The author declares no conflict of interest.

References

- Atienza D, Saiz E, Skovgaard A, Trepal I, Calbet A (2008). Life history and population dynamics of the marine cladoceran *Penilia avirostris* (Branchiopoda: Cladocera) in the Catalan Sea (NW Mediterranean). *Journal of Plankton Research* 30 (4): 345-357. <https://doi.org/10.1093/plankt/fbm109>
- Bedikoğlu D, Yılmaz IN, Demirel N (2022). Reproductive strategies and population characteristics of key Cladoceran species in the Sea of Marmara. *Regional Studies in Marine Science* 54: 102450. <https://doi.org/10.1016/j.rsma.2022.102450>
- Bişinicu E, Lazăr L, Timofte F (2023). Dynamics of zooplankton along the Romanian Black Sea coastline: temporal variation, community structure, and environmental drivers. *Diversity* 15 (9): 1024. <https://doi.org/10.3390/d15091024>
- Bryan BB (1979). The diurnal reproductive cycle of *Evadne tergestina* Claus (Cladocera, Podonidae) in Chesapeake Bay, U.S.A. *Crustaceana* 36 (3): 229-236. <https://doi.org/10.1163/156854079X00708>
- Buyukates Y, İnanmaz ÖE (2007). Temporal variations in vertical distribution and occurrence of marine cladocerans in an urbanized harbour, Dardanelles, Turkey. *Crustaceana* 80 (11): 1293-1302. <https://doi.org/10.1163/156854007782605583>
- Buyukates Y, İnanmaz ÖE (2009). Cladocerans of urbanized harbour: effects of environmental parameters on vertical distribution, occurrence, abundance and seasonal variation. *Crustaceana* 82 (5): 543-554. <https://doi.org/10.1163/156854009X407669>
- Cházaro-Olvera S, Winfield I, Ortiz M, Torres-Cabrera DA (2024). Population dynamics of *Penilia avirostris* and *Pseudevadne tergestina* (Diplostraca) in the Veracruz Reef System National Park, southwest Gulf of Mexico. *Latin American Journal of Aquatic Research* 52 (3): 486-500. <https://doi.org/10.3856/vol52-issue3-fulltext-3184>
- Della Croce N, Angelino M (1987). Marine Cladocera in the Gulf of Mexico and the Caribbean Sea. *Cahiers de Biologie Marine* 28 (2): 263-268. <https://doi.org/10.21411/CBM.A.74DED839>
- Du P, Ye W-J, Deng B-P, Mao M, Zhu YL et al. (2022). Long-term changes in zooplankton in the Changjiang estuary from the 1960s to 2020. *Frontiers in Marine Science* 9: 961591. <https://doi.org/10.3389/fmars.2022.961591>
- Egloff DA, Fofonoff PW, Onbé T (1997). Reproductive biology of marine cladocerans. *Advances in Marine Biology* 31: 79-167. [https://doi.org/10.1016/S0065-2881\(08\)60222-9](https://doi.org/10.1016/S0065-2881(08)60222-9)
- Forró L, Korovchinsky NM, Kotov AA, Petrusek A (2008). Global diversity of diplostraceans (Cladocera; Crustacea) in freshwater. *Hydrobiologia* 595: 177-184. <https://doi.org/10.1007/s10750-007-9013-5>
- Graeb BDS, Dettmers JM, Wahl DH, Cáceres CE (2004). Fish size and prey availability affect growth, survival, prey selection, and foraging behavior of larval yellow perch. *Transactions of the American Fisheries Society* 133 (3): 504-514. <https://doi.org/10.1577/T03-050.1>
- Gülşahin N, Tarkan AN (2012). Seasonal changes in distribution and abundance of the cladoceran species in relation to environmental factors in Gökova Bay (Mugla, Aegean Sea, Turkey). *Fresenius Environmental Bulletin* 21 (7a): 1853-1863.
- Hairston NG, Cáceres CE (1996). Distribution of crustacean diapauses: micro- and macroevolutionary pattern and process. *Hydrobiologia* 320 (1-3): 27-44. <https://doi.org/10.1007/BF00016802>
- He X, Lei M, Cheng F, Hu S (2023). *In situ* food compositions reveal niche partitioning in small marine cladocerans and copepods in Daya Bay, South China Sea. *Marine Ecology Progress Series* 716: 47-61. <https://doi.org/10.3354/meps14363>
- Hebert PDN, Ward RD (1972). Inheritance during parthenogenesis in *Daphnia magna*. *Genetics* 71 (4): 639-642. <https://doi.org/10.1093/genetics/71.4.639>
- Isinibilir M (2009). Annual crustacean zooplankton succession (Copepoda and Cladocera) in the upper layer of Ahirkapi coastal waters (northeastern Sea of Marmara). *Crustaceana* 82 (6): 669-678. <https://doi.org/10.1163/156854008x380273>
- Killi N (2020). Spatio-temporal variation in the distribution and abundance of marine cladocerans in relation to environmental factors in a productive lagoon (Güllük Bay, SW Aegean Sea, Turkey). *Oceanological and Hydrobiological Studies* 49 (4): 374-382. <https://doi.org/10.1515/ohs-2020-0032>
- Kodama T, Ohshimo S, Tanaka H, Ashida H, Kameda T et al. (2021). Abundance and habitats of marine cladocerans in the Sea of Japan over two decades. *Progress in Oceanography* 194: 102561. <https://doi.org/10.1016/j.pocean.2021.102561>
- Lankov A, Ojaveer H, Simm M, Pöllupüü M, Möllmann C (2010). Feeding ecology of pelagic fish species in the Gulf of Riga (Baltic Sea): the importance of changes in the zooplankton community. *Journal of Fish Biology* 77 (10): 2268-2284. <https://doi.org/10.1111/j.1095-8649.2010.02805.x>
- Lebedeva LP, Lukasheva TA, Anokhina LL, Chasovnikov VK (2015). Interannual variability in the zooplankton community in Golubaya Bay (Northeastern part of the Black Sea) in 2002-2012. *Oceanology* 55 (3): 355-363. <https://doi.org/10.1134/S0001437015030091>
- Marazzo A, Valentin JL (2003a). *Penilia avirostris* (Crustacea, Ctenopoda) in a tropical bay: variations in density and aspects of reproduction. *Acta Oecologica* 24 (Supplement 1): S251-S257. [https://doi.org/10.1016/S1146-609X\(03\)00019-5](https://doi.org/10.1016/S1146-609X(03)00019-5)
- Marazzo A, Valentin JL (2003b). Population parameters of *Pleopis polyphemoides* (Crustacea, Cladocera) in a tropical bay. *Estuarine, Coastal Shelf Science* 57 (5-6): 1015-1021. [http://dx.doi.org/10.1016/S0272-7714\(03\)00007-6](http://dx.doi.org/10.1016/S0272-7714(03)00007-6)
- Marazzo A, Valentin JL (2004a). Population dynamics of *Pseudevadne tergestina* (Branchiopoda: Onychopoda) in Guanabara Bay, Brazil. *Brazilian Archives of Biology and Technology* 47 (5): 713-723. <https://doi.org/10.1590/S1516-89132004000500006>

- Atienza D, Saiz E, Skovgaard A, Trepas I, Calbet A (2008). Life history and population dynamics of the marine cladoceran *Penilia avirostris* (Branchiopoda: Cladocera) in the Catalan Sea (NW Mediterranean). *Journal of Plankton Research* 30 (4): 345-357. <https://doi.org/10.1093/plankt/fbm109>
- Bedikoğlu D, Yılmaz İN, Demirel N (2022). Reproductive strategies and population characteristics of key Cladoceran species in the Sea of Marmara. *Regional Studies in Marine Science* 54: 102450. <https://doi.org/10.1016/j.rsma.2022.102450>
- Bişinicu E, Lazăr L, Timofte F (2023). Dynamics of zooplankton along the Romanian Black Sea coastline: temporal variation, community structure, and environmental drivers. *Diversity* 15 (9): 1024. <https://doi.org/10.3390/d15091024>
- Bryan BB (1979). The diurnal reproductive cycle of *Evadne tergestina* Claus (Cladocera, Podonidae) in Chesapeake Bay, U.S.A. *Crustaceana* 36 (3): 229-236. <https://doi.org/10.1163/156854079X00708>
- Buyukates Y, İnanmaz ÖE (2007). Temporal variations in vertical distribution and occurrence of marine cladocerans in an urbanized harbour, Dardanelles, Turkey. *Crustaceana* 80 (11): 1293-1302. <https://doi.org/10.1163/156854007782605583>
- Buyukates Y, İnanmaz ÖE (2009). Cladocerans of urbanized harbour: effects of environmental parameters on vertical distribution, occurrence, abundance and seasonal variation. *Crustaceana* 82 (5): 543-554. <https://doi.org/10.1163/156854009X407669>
- Cházaro-Olvera S, Winfield I, Ortiz M, Torres-Cabrera DA (2024). Population dynamics of *Penilia avirostris* and *Pseudevadne tergestina* (Diplostraca) in the Veracruz Reef System National Park, southwest Gulf of Mexico. *Latin American Journal of Aquatic Research* 52 (3): 486-500. <https://doi.org/10.3856/vol52-issue3-fulltext-3184>
- Della Croce N, Angelino M (1987). Marine Cladocera in the Gulf of Mexico and the Caribbean Sea. *Cahiers de Biologie Marine* 28 (2): 263-268. <https://doi.org/10.21411/CBM.A.74DED839>
- Du P, Ye W-J, Deng B-P, Mao M, Zhu YL et al. (2022). Long-term changes in zooplankton in the Changjiang estuary from the 1960s to 2020. *Frontiers in Marine Science* 9: 961591. <https://doi.org/10.3389/fmars.2022.961591>
- Egloff DA, Fofonoff PW, Onbé T (1997). Reproductive biology of marine cladocerans. *Advances in Marine Biology* 31: 79-167. [https://doi.org/10.1016/S0065-2881\(08\)60222-9](https://doi.org/10.1016/S0065-2881(08)60222-9)
- Forró L, Korovchinsky NM, Kotov AA, Petrusek A (2008). Global diversity of diplostraceans (Cladocera; Crustacea) in freshwater. *Hydrobiologia* 595: 177-184. <https://doi.org/10.1007/s10750-007-9013-5>
- Graeb BDS, Dettmers JM, Wahl DH, Cáceres CE (2004). Fish size and prey availability affect growth, survival, prey selection, and foraging behavior of larval yellow perch. *Transactions of the American Fisheries Society* 133 (3): 504-514. <https://doi.org/10.1577/T03-050.1>
- Gülşahin N, Tarkan AN (2012). Seasonal changes in distribution and abundance of the cladoceran species in relation to environmental factors in Gökova Bay (Mugla, Aegean Sea, Turkey). *Fresenius Environmental Bulletin* 21 (7a): 1853-1863.
- Hairston NG, Cáceres CE (1996). Distribution of crustacean diapauses: micro- and macroevolutionary pattern and process. *Hydrobiologia* 320 (1-3): 27-44. <https://doi.org/10.1007/BF00016802>
- He X, Lei M, Cheng F, Hu S (2023). *In situ* food compositions reveal niche partitioning in small marine cladocerans and copepods in Daya Bay, South China Sea. *Marine Ecology Progress Series* 716: 47-61. <https://doi.org/10.3354/meps14363>
- Hebert PDN, Ward RD (1972). Inheritance during parthenogenesis in *Daphnia magna*. *Genetics* 71 (4): 639-642. <https://doi.org/10.1093/genetics/71.4.639>
- Isinibilir M (2009). Annual crustacean zooplankton succession (Copepoda and Cladocera) in the upper layer of Ahirkapi coastal waters (northeastern Sea of Marmara). *Crustaceana* 82 (6): 669-678. <https://doi.org/10.1163/156854008x380273>
- Killi N (2020). Spatio-temporal variation in the distribution and abundance of marine cladocerans in relation to environmental factors in a productive lagoon (Güllük Bay, SW Aegean Sea, Turkey). *Oceanological and Hydrobiological Studies* 49 (4): 374-382. <https://doi.org/10.1515/ohs-2020-0032>
- Kodama T, Ohshimo S, Tanaka H, Ashida H, Kameda T et al. (2021). Abundance and habitats of marine cladocerans in the Sea of Japan over two decades. *Progress in Oceanography* 194: 102561. <https://doi.org/10.1016/j.pocean.2021.102561>
- Lankov A, Ojaveer H, Simm M, Pöllupüü M, Möllmann C (2010). Feeding ecology of pelagic fish species in the Gulf of Riga (Baltic Sea): the importance of changes in the zooplankton community. *Journal of Fish Biology* 77 (10): 2268-2284. <https://doi.org/10.1111/j.1095-8649.2010.02805.x>
- Lebedeva LP, Lukasheva TA, Anokhina LL, Chasovnikov VK (2015). Interannual variability in the zooplankton community in Golubaya Bay (Northeastern part of the Black Sea) in 2002-2012. *Oceanology* 55 (3): 355-363. <https://doi.org/10.1134/S0001437015030091>
- Marazzo A, Valentin JL (2003a). *Penilia avirostris* (Crustacea, Ctenopoda) in a tropical bay: variations in density and aspects of reproduction. *Acta Oecologica* 24 (Supplement 1): S251-S257. [https://doi.org/10.1016/S1146-609X\(03\)00019-5](https://doi.org/10.1016/S1146-609X(03)00019-5)
- Marazzo A, Valentin JL (2003b). Population parameters of *Pleopis polyphemoides* (Crustacea, Cladocera) in a tropical bay. *Estuarine, Coastal Shelf Science* 57 (5-6): 1015-1021. [http://dx.doi.org/10.1016/S0272-7714\(03\)00007-6](http://dx.doi.org/10.1016/S0272-7714(03)00007-6)
- Marazzo A, Valentin JL (2004a). Population dynamics of *Pseudevadne tergestina* (Branchiopoda: Onychopoda) in Guanabara Bay, Brazil. *Brazilian Archives of Biology and Technology* 47 (5): 713-723. <https://doi.org/10.1590/S1516-89132004000500006>
- Marazzo A, Valentin JL (2004b). Reproductive aspects of marine cladocerans *Penilia avirostris* and *Pseudevadne tergestina* (Crustacea, Branchiopoda) in the outer part of Guanabara Bay, Brazil. *Brazilian Journal of Biology* 64 (3A): 543-549. <https://doi.org/10.1590/S1519-69842004000300017>

- Mayer CM, Wahl DH (1997). The relationship between prey selectivity and growth and survival in a larval fish. *Canadian Journal of Fisheries and Aquatic Sciences* 54 (7): 1504-1512. <https://cdnsiencepub.com/doi/abs/10.1139/f97-056>
- Miyashita LK, Gaeta SA, Lopes RM (2011). Life cycle and reproductive traits of marine podonids (Cladocera, Onychopoda) in a coastal subtropical area. *Journal of Plankton Research* 33 (5): 779-792. <https://doi.org/10.1093/plankt/fbq147>
- Miyashita LK, Pompeu M, Gaeta SA, Lopes RM (2010). Seasonal contrasts in abundance and reproductive parameters of *Penilia avirostris* (Cladocera, Ctenopoda) in a coastal subtropical area. *Marine Biology* 157 (11): 2511-2519. <http://dx.doi.org/10.1007/s00227-010-1515-4>
- Mullin MM, Onbé T (1992). Diel reproduction and vertical distributions of the marine cladocerans, *Evadne tergestina* and *Penilia avirostris*, in contrasting coastal environments. *Journal of Plankton Research* 14 (1): 41-59. <http://dx.doi.org/10.1093/plankt/14.1.41>
- Oghenekaro EU, Chigbu P (2019). Population dynamics and life history of marine Cladocera in the Maryland Coastal Bays, USA. *Journal of Coastal Research* 35 (6): 1225-1236. <https://doi.org/10.2112/JCOASTRES-D-18-00062.1>
- Onbé T (1985). Seasonal fluctuations in the abundance of populations of marine cladocerans and their resting eggs in the Inland Sea of Japan. *Marine Biology* 87 (1): 83-88. <http://dx.doi.org/10.1007/BF00397009>
- Onbé T (1999). Ctenopoda and Onychopoda (Cladocera). In: Boltovskoy D (editor). *South Atlantic Zooplankton*. Leiden, Netherlands: Backhuys Publishers, pp. 797-813.
- Öztekin A, Üstün F, Bat L, Tabak A (2024). Microplastic contamination of the seawater in the Hamsilos Bay of the Southern Black Sea. *Water, Air, & Soil Pollution* 235 (6): 325. <https://doi.org/10.1007/s11270-024-07138-w>
- Parsons TR, Maita Y, Lalli CM (1984). *A Manual of Chemical and Biological Methods for Seawater Analysis*. Oxford, UK: Pergamon Press.
- Pavlova EV (1959). Development cycle and some data on the growth of *Penilia avirostris* Dana in Sevastopol Bay. *Trudy Sevastopol'skoi Biologicheskoi Stantsii* 11: 54-62 (in Russian).
- Platt T, Yamamura N (1986). Prenatal mortality in a marine cladoceran, *Evadne nordmanni*. *Marine Ecology Progress Series* 29: 127-139.
- Postel L, Fock H, Hagen W (2000). Collecting zooplankton. In: Harris RP, Wiebe PH, Lenz J, Skjoldal HR, Huntley M (editors). *ICES Zooplankton Methodology Manual*. London, UK: Academic Press, pp. 55-81.
- Pöllupüü M, Simm M, Ojaveer H (2010). Life history and population dynamics of the marine cladoceran *Pleopis polyphemoides* (Leuckart) (Cladocera, Crustacea) in a shallow temperate Pärnu Bay (Baltic Sea). *Journal of Plankton Research* 32 (10): 1459-1469. <https://doi.org/10.1093/plankt/fbq063>
- Ramírez FC, Pérez Seijas GM (1985). New data on the ecological distribution of cladocerans and first local observations on reproduction of *Evadne nordmanni* and *Podon intermedius* (Crustacea, Cladocera) in Argentine sea waters. *Physis, Section A: Natural Sciences* 43 (105): 131-143.
- Sabatés A, Zaragoza N, Raya V (2015). Distribution and feeding dynamics of larval red mullet (*Mullus barbatus*) in the NW Mediterranean: the important role of cladocera. *Journal of Plankton Research* 37 (4): 820-833. <https://doi.org/10.1093/plankt/fbv040>
- Schroeder A, Camatti E, Pansera M, Pallavicini A (2022). Applying DNA metabarcoding for the diet investigation of the invasive ctenophore *Mnemiopsis leidyi* in a transitional environment. *Research Square* (preprint). <https://doi.org/10.21203/rs.3.rs-1307217/v1>
- Tang KW, Chen QC, Wong CK (1995). Distribution and biology of marine cladocerans in the coastal waters of southern China. *Hydrobiologia* 307 (1-3): 99-107. <http://dx.doi.org/10.1007/BF00032001>
- Terbiyik Kurt T, Polat S (2017). Introduction of a new Indo-Pacific marine cladoceran to the Mediterranean Sea. *Mediterranean Marine Science* 18 (3): 517-523. <https://doi.org/10.12681/mms.13885>
- Terbiyik Kurt T, Toklu Aliçlı B, Polat S (2018). A spatial and temporal distribution of marine cladoceran species in the surface waters of Iskenderun Bay. *Aquatic Research* 1 (2): 77-85. <https://doi.org/10.3153/AR18009>
- Terbiyik Kurt T, Polat S (2014). Characterization of seasonal and inter-annual changes in the abundance of species of marine Cladocera on the Turkish coast of the northeastern Levantine Basin. *Crustaceana* 87 (7): 769-783. <https://doi.org/10.1163/15685403-00003325>
- Toklu Aliçlı B, Balkis N, Balci M (2014). Seasonal distribution of zooplankton in the Gulf of Erdek (The Marmara Sea) and the impact of ecological variables. *Fresenius Environmental Bulletin* 23 (12): 3013-3021.
- Uye S, Liang D (2022). Seasonal population dynamics, production, and feeding of the chaetognath *Aidanosagitta crassa* in a temperate eutrophic inlet. *Plankton and Benthos Research* 17 (3): 312-326. <https://doi.org/10.3800/pbr.17.312>
- Uygun O, Hoşsucu B (2024). Temporal dynamics of the ichthyoplankton assemblages in the central Aegean Sea during one year and the effects of ecological factors. *Regional Studies in Marine Science* 72: 103439. <https://doi.org/10.1016/j.rsma.2024.103439>
- Uygun O, Üstün F (2024). Ichthyoplankton assemblages from the coasts of Hamsilos Nature Park, Sinop, Southern Black Sea: biodiversity, abundance, and relationships with environmental variables. *Water* 16 (18): 2670. <https://doi.org/10.3390/w16182670>
- Üstün F (2019). Seasonal cycle of zooplankton abundance and biomass in Hamsilos Bay, Sinop, Southern Black Sea, Turkey. *Journal of Natural History* 53 (7-8): 365-389. <https://doi.org/10.1080/00222933.2019.1592257>

- Üstün F, Bat L, Mutlu E (2018). Seasonal variation and taxonomic composition of mesozooplankton in the Southern Black Sea (off Sinop) between 2005 and 2009. Turkish Journal of Zoology 42 (5): 541-556. <https://doi.org/10.3906/zoo-1801-13>
- Üstün F, Bat L, Şahin F, Birinci Özdemir Z, Kideys AE (2016). Seasonality of mesozooplankton in the Southern Black Sea (off Sinop) between 2002 and 2004. Journal of New Results in Science 5 (11): 87-101.
- Wong CK, Ji C, Nip THM (2004). Diel cycle in the percentage abundance of parthenogenetic females with embryos of different developmental stages in four species of marine cladocerans. Journal of Plankton Research 26 (9): 1095-1103. <https://doi.org/10.1093/plankt/fbh102>
- Wong CK, Li VCY, Chan A (2008). Diel cycles of reproduction and vertical migration in the marine cladocerans *Pseudevadne tergestina* and *Penilia avirostris*. Journal of Plankton Research 30 (1): 65-73. <https://doi.org/10.1093/plankt/fbm096>
- Yildiz İ (2025). Cladoceran species and their population abundance across the Black Sea from west to east during summer 2020. Journal of Natural History 59 (5-8): 271-283. <https://doi.org/10.1080/00222933.2024.2448188>